

1 **Metabolomic Changes Following GenX and PFBS Exposure in Developing**
2 **Zebrafish**

3
4 **Authors:** Fiona Dunn,¹ Shannon E. Paquette,² Kurt D. Pennell,¹ Jessica S. Plavicki,^{2*} Katherine
5 E. Manz^{1,3*}

6 ¹School of Engineering, Brown University, 184 Hope Street, Providence, RI, 02912

7 ²Department of Pathology and Laboratory Medicine, Brown University, 70 Ship Street,
8 Providence, RI, 02903

9 ³Department of Environmental Health Sciences, University of Michigan, 1415 Washington
10 Heights, Ann Arbor, MI 48109

11 *Email: jessica_plavicki@brown.edu; katmanz@umich.edu

12 **Keywords:** GenX, PFBS, Zebrafish, Embryos, Metabolomics, Body Burden

24 **ABSTRACT**

25 Short chain per- and polyfluoroalkyl substances (PFAS), including hexafluoropropylene oxide
26 dimer acid (GenX) and perfluorobutane sulfonate (PFBS), are replacement chemicals for
27 environmentally persistent, long-chain PFAS. GenX and PFBS have been detected in surface and
28 groundwater worldwide; however, few studies provide information on the metabolic effects or
29 risks associated with their exposures. In this study, larval zebrafish were used as a model for
30 investigating the toxicity of early-life exposure to either GenX or PFBS. Zebrafish were
31 chronically exposed from 4 hours post-fertilization (hpf) to 6 days post-fertilization (dpf) to 150
32 μM GenX or 95.0 μM PFBS. Ultra-high-performance liquid chromatography paired with high
33 resolution mass spectrometry was used to quantify uptake of PFAS into zebrafish larvae and
34 perform targeted and untargeted metabolomics. Our results indicate that PFBS was 20.4% more
35 readily absorbed into the zebrafish larvae compared to GenX. Additionally, PFBS exposure
36 significantly altered 13 targeted metabolites and 21 metabolic pathways, while GenX exposure
37 significantly altered 1 targeted metabolite and 17 metabolic pathways. In both the GenX and PFBS
38 exposures, the greatest number of altered metabolic pathways occurred in the global amino acid
39 pathways, including alterations in the carbohydrate, lipid, cofactors and vitamins, nucleotide, and
40 xenobiotics metabolisms. Our results indicate that GenX and PFBS impact the zebrafish
41 metabolome, with implications of altered growth and development and global metabolic
42 dysregulation.

43

44

45

46

47 **1. INTRODUCTION**

48 Per- and polyfluoroalkyl substances (PFAS) have emerged as contaminants of concern due
49 to their bioaccumulative properties and potential adverse effects on animal and human health.
50 PFAS are widely used in a range of consumer products, including non-stick or stain/water-resistant
51 products (Sunderland et al., 2019). PFAS are a broad class of chemicals containing at least one
52 fully fluorinated carbon atom and a functional head group (Rayne and Forest, 2009; Wallington et
53 al., 2021). Due to the strength of the carbon fluorine bond, PFAS are recalcitrant to most chemical
54 and biological degradation processes and persist in the environment (Glüge et al., 2020; Kucharzyk
55 et al., 2017). Exposure to PFAS is associated with adverse effects to the immune, reproductive,
56 and endocrine systems and diseases, including cancers in both animal models and humans (Barbo
57 et al., 2023; Podder et al., 2021; Sunderland et al., 2019).

58 Due to concerns about the potential health risks of PFAS exposure, long-chain PFAS
59 (containing eight or more carbons in the perfluoroalkyl chain, including legacy perfluorooctanoic
60 acid (PFOA) and perfluorooctane sulfonate (PFOS)) were voluntarily phased out of production in
61 the early 2000s, resulting in increased production of replacement short-chain PFAS (containing
62 less than seven carbons) (Janousek et al., 2019; Kabadi et al., 2020). Because of their shorter
63 carbon chain length, these replacement chemicals were intended to be less persistent in the
64 environment and are now used in industrial applications across the world (Bao et al., 2018; Wang
65 et al., 2016). In 2022, the US Environmental Protection Agency (EPA) set health advisory levels
66 for four PFAS compounds in drinking water: PFOA and PFOS (C₈) (0.004 ppt and 0.02 ppt,
67 respectively) and their short chain replacement chemicals GenX (hexafluoropropylene oxide dimer
68 acid (HFPO-DA)) (C₆) and perfluorobutane sulfonate (PFBS) (C₄) (10 and 2,000 ppt, respectively)
69 (USEPA, 2022). In March of 2023, the EPA proposed maximum contaminant level (MCL) goals

70 for PFOA and PFOS in drinking water at 4 ppt each (USEPA, 2023). The MCL also proposed a
71 Hazard Index, which sums the ratios of the water concentration to the level determined not to cause
72 health effects for two short chain PFAS - GenX and PFBS – and two long chain PFAS -
73 perfluorononanoic acid (PFNA) and perfluorohexane sulfonate (PFHxS) (USEPA, 2022; USEPA,
74 2023).

75 GenX and PFBS have been detected in surface and ground waters in many countries, with
76 demonstrated bioaccumulation in nearby plants, animals, and humans (Blake et al., 2020; Taniyasu
77 et al., 2003). The potential health effects of these replacement PFAS are largely unknown, but
78 preliminary animal studies suggest that there may be an association between GenX and PFBS
79 exposure and alterations to thyroid function, reproductive problems, low birth weight, and growth
80 problems (Blake et al., 2020; Chen et al., 2018; Newsted et al., 2008). Zebrafish are widely used
81 in toxicology research to study the effects of xenobiotic exposure on the vertebrate exposome,
82 which reflects the biological outcomes of exposures (Howe et al., 2013). Due to their rapid ex-
83 utero embryonic development and genomic similarity to humans, zebrafish are an excellent model
84 organism for studying vertebrate embryonic development and health (Gaballah et al., 2020; Howe
85 et al., 2013). Previous studies of early-life exposure of zebrafish to short-chain PFAS reveal
86 alterations to exposed zebrafish such as abnormal behavior, altered growth and development,
87 cardiac edema, and developmental neurotoxicity (Gebreab et al., 2020; Min et al., 2023; Rericha
88 et al., 2022). As short-chain replacers for PFAS become more prevalent in the environment, it is
89 essential to study biological systems that may be affected by exposure to these chemicals.

90 Incorporating “omics”, such as genomics (the study of gene expression), transcriptomics
91 (the study of tRNA), proteomics (the study of proteins), and metabolomics (the study of
92 metabolites and pathways), into zebrafish studies provides molecular insight into physiological

93 alterations caused by chemical exposures (da Silva et al., 2021; Sukardi et al., 2010). Since
94 metabolites are products of cellular processes, metabolic alterations directly indicate physiological
95 and phenotypic alterations (da Silva et al., 2021; Lai et al., 2021). When applied to embryonic or
96 larval zebrafish, metabolomics provides important insight into the effects of exposures on growth
97 and development and molecular mechanisms for chemically induced toxicological endpoints (da
98 Silva et al., 2021; Kossack et al., 2023).

99 Previous metabolomic studies of zebrafish exposure to the legacy long-chain PFAS,
100 primarily PFOA and PFOS, have demonstrated alterations to the lipid, carbohydrate, and amino
101 acid metabolisms (Cheng et al., 2016; Du et al., 2016; Gebreab et al., 2020). These studies included
102 short-term and chronic exposure during several different life stages of zebrafish (embryos, larvae,
103 and adults) and compared metabolomic effects of PFOA, PFOS, and next-generation PFAS
104 (GenX, PFBS, PFO3TDA (perfluoro-3,6,9-trioxadecanoic acid)). One study exposed zebrafish
105 embryos to either GenX or PFOA at concentrations up to 100 mg L⁻¹ and compared the embryonic
106 toxicity and metabolic alterations induced by exposure to PFOA or GenX, finding that exposures
107 to either chemical altered metabolites that are essential to liver function and development (Gebreab
108 et al., 2020). A second study of embryonic exposure to 10 mg L⁻¹ PFBS indicated an induced stress
109 response, neurotoxicity, and altered folate biosynthesis and methionine metabolisms (Hu et al.,
110 2022). Both studies used very different exposure doses and examined different resulting metabolic
111 impacts, thus knowledge of short-chain PFAS uptake and the use of similar exposure levels are
112 needed to assess metabolomic alterations.

113 The objective of this study was to understand global metabolic alterations in developing
114 zebrafish resulting from exposure to the short-chain PFAS GenX and PFBS. Using targeted
115 analysis, GenX and PFBS uptake and body burden of exposed zebrafish larvae were quantified

116 and targeted metabolomics was performed to quantify levels of 33 amino acid metabolites to
117 compare between exposed and non-exposed larvae. Using untargeted metabolomics, global
118 metabolic alterations induced by GenX and PFBS exposure were discovered, which provided
119 insight into the molecular mechanisms of toxicity in early-life zebrafish exposure to short-chain
120 PFAS. The results of this study provide new information on short-chain PFAS uptake and
121 metabolic alterations in exposed zebrafish larvae.

122

123 **2. MATERIALS AND METHODS**

124 ***2.1 Chemicals***

125 All solutions were prepared using water purified by a Millipore Milli-Q® Reference
126 purification system (18.2 MΩ.cm at 25°C and total organic content below 5 ppb). GenX
127 (Ammonium perfluoro (2-methyl-3-oxahexanote), 95% purity) used for the exposure solutions
128 was purchased from Manchester Organics (Runcorn, England). Dimethyl sulfoxide (DMSO) and
129 PFBS (perfluorobutane sulfonate) for exposure solutions were obtained from MilliporeSigma (St.
130 Louis, MO). Isotopically labeled perfluorinated compound internal standards for quantitation of
131 GenX and PFBS were purchased from Wellington Laboratories (Guelph, Ontario, Canada). All
132 solvents, including Ultra High-Performance Liquid Chromatography (UHPLC) grade acetonitrile
133 and water (99.9%), were purchased from Fisher Scientific (Waltham, MA). Isotopically labeled
134 metabolites (containing L-alanine, L-leucine, L-phenylalanine, L-tryptophan, and L-tyrosine)
135 were obtained from Cambridge Isotopes (MSK-QC-KIT, Tewksbury, MA). Amino acid
136 metabolite standards (A9906-10 mL) and other targeted metabolites (DL-3-Ureidoisobutyric acid,
137 N-Acetylputrescine hydrochloride, γ -Aminobutyric acid, L-Asparagine, L-Pipecolic acid, L-
138 Glutamine, 3-Ureidopropionic acid) were purchased from Sigma Aldrich (St. Louis, MO).

139

140 **2.2 Zebrafish Spawning**

141 All maintenance and experimental procedures involving zebrafish (*Danio rerio*) were
142 approved by the Brown University Institutional Animal Care and Use Committee (IACUC; 19-12-
143 0003) and adhered to the National Institute of Health's "Guide for the Care and Use of Laboratory
144 Animals." The zebrafish were kept in an aquatic habitat (Aquaneering Inc., San Diego, CA) with
145 a 14:10 hour light-dark cycle (Westerfield, 2000) and water temperature control ($28.5 \pm 2^\circ\text{C}$),
146 filtration, purification, automatic pH and conductivity stabilization, and ultraviolet (UV)
147 irradiation disinfection.

148 The night before spawning, adult AB (wildtype) zebrafish were acclimated to their new
149 tank (1.7 L specialized, gridded spawning tanks (Techniplast, USA)) conditions to facilitate timed
150 reproduction. A transparent partition was used to separate sexes. The tank partition was removed
151 within 2 hours of light cycle onset so that the zebrafish could spawn for 1 hour. Three independent
152 spawning events were performed on separate days to acquire three biological replicates of each
153 dosing experiment. Spawning occurred with separate parents on three different dates for each
154 exposure compound (GenX, PFBS, or vehicle (egg water containing 0.1% DMSO) to ensure an
155 adequate number and varied parenthood of embryos. All embryos were collected in 100 mm non-
156 treated culture petri dishes containing fresh egg water (60 mg L⁻¹ Instant Ocean Sea Salts;
157 Aquarium Systems, Mentor, OH) after spawning. At 4 hpf, any embryos that did not show adequate
158 cellular development by 4 manual screening for quality were discarded.

159 **2.3 Exposure**

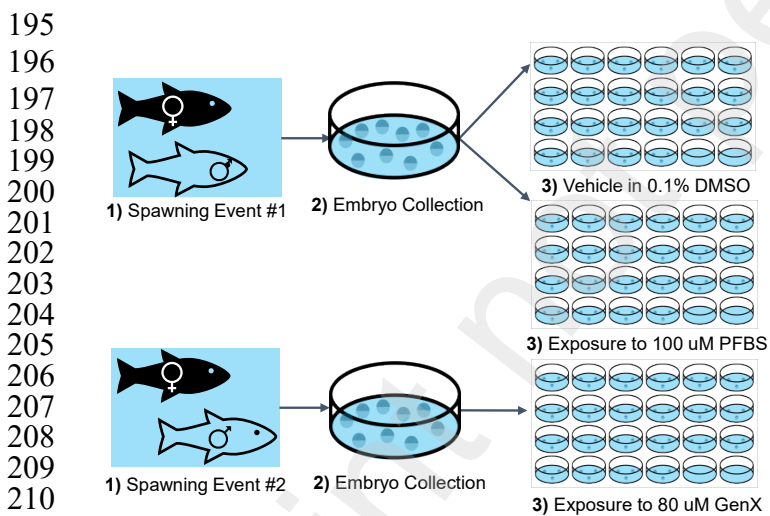
160 *Exposure Solutions:* Aqueous stock solutions of PFBS (64 mM) and GenX (41.48 mM)
161 were prepared and validated at the start of the study. Two exposure solutions were prepared by

162 diluting the stock solutions in fresh egg water in polypropylene tubes to obtain desired exposure
163 concentrations of 150 μM (49.5 mg L^{-1}) for GenX and 100 μM (30.0 mg L^{-1}) for PFBS. Solutions
164 were validated using targeted mass spectrometry (further described below) and the actual exposure
165 solution concentrations were 150 μM (49.5 mg L^{-1}) for GenX and 95.0 μM (28.5 mg L^{-1}) for PFBS.
166 For GenX, a previously published concentration between 100 and 200 μM demonstrated no
167 physiological deformities or adverse effects in larval zebrafish at our desired timepoints (Gebreab
168 et al., 2020); indeed, no physiological deformities or excess mortality were noted in our study
169 following exposure to 150 μM GenX by 6 days post-fertilization (dpf). Likewise, the concentration
170 selected for PFBS was also based on previously published work by Gaballah et al., who found that
171 at 100 μM PFBS, exposed zebrafish had no physiological deformities or differences in locomotor
172 activity in comparison with non-exposed groups (Gaballah et al., 2020). These exposure
173 concentrations were verified not to produce physical deformities indicative of poor health, such as
174 a bent spine or cardiac edema, by performing daily observations throughout the exposure period
175 using a microscope.

176 *Fish Exposure:* At 4 hpf, we plated 3 fertilized embryos per well in three sterile 24-well
177 plates (ThermoFisher, Waltham, MA). Each well was dosed with 2 mL of either GenX, PFBS, or
178 egg water with 0.1% DMSO (vehicle control) (**Figure 1**). DMSO was chosen as the vehicle
179 because PFBS was dissolved in DMSO to generate a stock solution. GenX is unstable in DMSO
180 and soluble in water (Gaballah et al., 2020); therefore, we dissolved GenX in purified deionized
181 water. A previous study compared exposure solutions containing PFBS dissolved in DMSO and
182 GenX in deionized water at the same concentrations and found that exposures resulted in changes
183 of similar magnitudes (Gaballah et al., 2020). The plates were then incubated at $28.5 \pm 1^\circ\text{C}$

184 incubator (Powers Scientific Inc., Pipersville, PA) from 4 hpf until 6 dpf. Fish were chronically
185 exposed without replacement of the treatment solution.

186 All embryos were manually dechorionated at 24 hpf to ensure unprotected exposure to
187 toxicants throughout the protocol and placed back into the incubator until collection at 6 dpf (Henn
188 and Braunbeck, 2011). For each exposure, 6 dpf larvae were pooled into 6 polypropylene
189 centrifuge tubes (2 mL) of n=10 larvae. Excess exposure solution was discarded from the larval
190 pools, and then the pools were washed three times in laboratory-grade deionized water with all
191 excess water removed between each round of washing. Samples were snap frozen in liquid
192 nitrogen to be stored at -80°C until ready for extraction. In total, 6 pools of 10 larvae were collected
193 per biological replicate for each exposure group, resulting in 180 larvae per exposure group and
194 540 total larvae collected across all groups.



212 **Figure 1:** Experimental design for embryo spawning and exposure dosing for each exposure.

213 *Sample Extraction:* 30 μL of PFAS internal standard solution and 270 μL of acetonitrile containing
214 an isotopically labeled metabolite internal standard solution were added to each sample. Samples
215 were sonicated for 30 minutes until homogenized, and then all samples were centrifuged at 755 x
216 g for 10 minutes using a Savant HSC-10K High Speed Centrifuge. For metabolomics, 200 μL of
217 the supernatant was transferred into an amber LC/MS vial containing a 300 μL glass insert. For
218 PFAS analysis, the remaining 100 μL of supernatant was diluted in acetonitrile and spiked with
219 internal standard to comprise 10% of the total volume of the sample. These samples were
220 transferred into 300 μL amber glass insert LC/MS vials. All samples were stored at 5°C prior to
221 analysis.

222 **2.4 Analytical Methods**

223 *PFAS Analysis:* Concentrations of GenX and PFBS were confirmed in the following
224 solutions: (1) GenX and PFBS stock solutions, (2) diluted GenX and PFBS exposure solutions and
225 matching vehicle (egg water containing 0.1% DMSO), and (3) zebrafish larval extracts. To
226 quantify PFBS or GenX in each solution, samples were diluted with UHPLC-grade acetonitrile
227 containing 10% isotope labeled internal standard.

228 GenX and PFBS concentrations were quantified based on a modified version of EPA Draft
229 Method 1633 using a high resolution Thermo QExactive HF-X Orbitrap MS equipped with a
230 Vanquish ultra-high-performance liquid chromatograph (UHPLC-Orbitrap-HRMS) (USEPA,
231 2021). Analyte separation was achieved by injecting 20 μL of each sample onto a Thermo Hypersil
232 Gold Vanquish C18 column (100 mm X 2.1 mm x 1.9 μm). Mobile phases consisted of 2 mM
233 ammonium acetate in 5% acetonitrile and 2 mM ammonium acetate in 100% acetonitrile. The MS
234 acquisitions were performed with negative electrospray ionization (ESI), in full scan data-
235 dependent (MS-ddMS²) with an inclusion list for GenX, PFBS, and internal standards, and with

236 100-1500 m/z scan range. An HCD collision cell filled with N₂ gas (produced by a Peak Scientific
237 Nitrogen Generator, Genius NM32LA) was used to perform MS² fragmentation. Additional
238 details on the chromatography, source settings, and MS data collection parameters are provided in
239 the SI (**Table S1**). PFAS concentrations were quantified using one quantification ion and eight-
240 point calibration curves with concentrations ranging from 0 to 20,000 ng/L for each compound.
241 Compound confirmation was performed using retention time and two confirming ions, when
242 available. The Limits of Detection (LODs) were 42.1 ng/L for GenX and 34.6 ng/L for PFBS,
243 which were determined from seven injections of calibration standards and multiplying the standard
244 deviation of these injections by 3.143 ($t_{0.99}$ at 6 degrees freedom) divided by the calibration curve
245 slope.

246 *Targeted and Untargeted Metabolomics:* Both targeted and untargeted metabolomics were
247 performed using the Thermo UHPLC-Orbitrap-HRMS. All samples were analyzed in triplicate
248 using a randomized order in a single batch. All samples were analyzed using two chromatography
249 methods, one with positive and negative ESI. For positive ESI (referred to as ESI+), a HILIC
250 column (Thermo Synchronis HILIC 100 mm X 2.1 mm x 3 μ m) was used. For negative ESI (referred
251 to as ESI-), a C18 column (Thermo Hypersil Gold Vanquish, 100 mm X 2.1 mm x 1.9 μ m) was
252 used. Additional details on the chromatography, source settings, and dd-MS² data collection
253 parameters are provided in the SI (**Table S2**). For targeted metabolomics detected in the ESI+
254 method, a ten-point calibration curve ranging from 1.25 to 333 nM was used to quantify the
255 metabolites in **Table S3**.

256 **2.5 Untargeted Metabolomics – Data Analysis**

257 XCalibur File Converter (ThermoFisher) was used to convert data files from .raw to .cdf
258 format. R packages apLCMS and xMSanalyzer were used to produce m/z feature tables (Denison

259 and Nagy, 2003; Uppal et al., 2013), which contained each feature (representative of a metabolite)
260 in a row and the m/z value, paired retention time (seconds), and unitless feature areas in columns.
261 All features containing greater than 20% non-detection were removed. Each feature was then log₂
262 transformed to normalize the data.

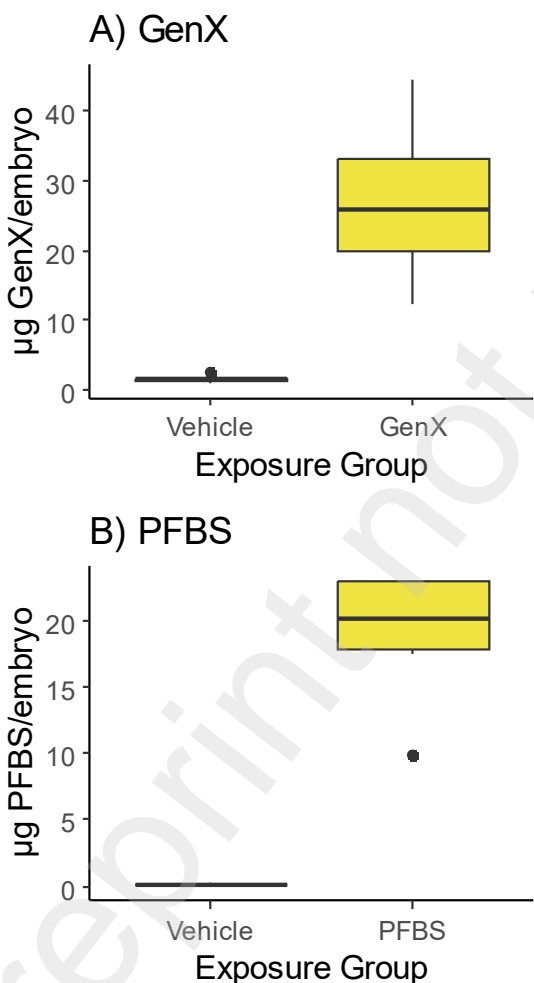
263 *Statistical Analysis:* An Analysis of Covariance (ANCOVA) test was performed for each
264 metabolomics data set (targeted and untargeted) for each exposure (GenX or PFBS) to test for
265 between-group variance vs. within-group variance and account for the varying birth dates and
266 parenthood of the zebrafish collected. For targeted data, ANCOVA results with an F-value > 0.05
267 were considered to be significantly altered metabolites (Mishra et al., 2019). In untargeted
268 metabolomics, a false discovery rate (FDR) of 20% was used to account for Type I errors in the
269 comparisons (Kossack et al., 2023) and any feature above this threshold was used for pathway
270 analysis.

271 *Pathway Analysis:* Pathway analysis was performed using MetaboAnalyst's Functional
272 Analysis for zebrafish (Danio rerio (KEGG)) (Chong and Xia, 2018). Features were matched to
273 metabolites with 5.0 ppm mass tolerance for ESI+ and ESI- data, including the feature F-value
274 (equivalent to p-value in an ANCOVA) obtained from statistical analysis and retention time. The
275 Mummichog algorithm was used with the default p-value cutoff of the top 10% of features, which
276 resulted in the following p-value (F-value) cut-offs: ESI+ GenX 0.25; ESI- GenX 0.1; ESI+ PFBS
277 0.05; ESI- PFBS 0.15 (Li et al., 2013). Our results include pathways that contained least 3
278 metabolites matched to features. Alterations to global metabolisms were mapped using the Kyoto
279 Encyclopedia of Genes and Genomes (KEGG) (Kanehisa and Goto, 2000).

280 **3. RESULTS**

281 ***3.1 Body Burden***

282 To establish uptake of PFBS and GenX, body burden per larva was determined for the six-
283 day exposure period. GenX body burden in the exposure and vehicle groups was 26.6 ± 10.8
284 $\mu\text{g}/\text{embryo}$ and $1.53 \pm 0.495 \mu\text{g}/\text{embryo}$, respectively (n=10, **Figure 2**). PFBS body burden in the
285 exposure and vehicle groups was $19.3 \pm 4.41 \mu\text{g}/\text{embryo}$ and $0.114 \pm 0.0413 \mu\text{g}/\text{embryo}$,
286 respectively (n=10, **Figure 2**). Based on the exposure concentration in each exposure well (n=3
287 embryos), the bioconcentration factors, or the ratio of the concentration inside the zebrafish to the
288 surrounding aqueous chemical concentration, were calculated to be 63.8 for PFBS and 50.6 for
289 GenX (**Table S4**).

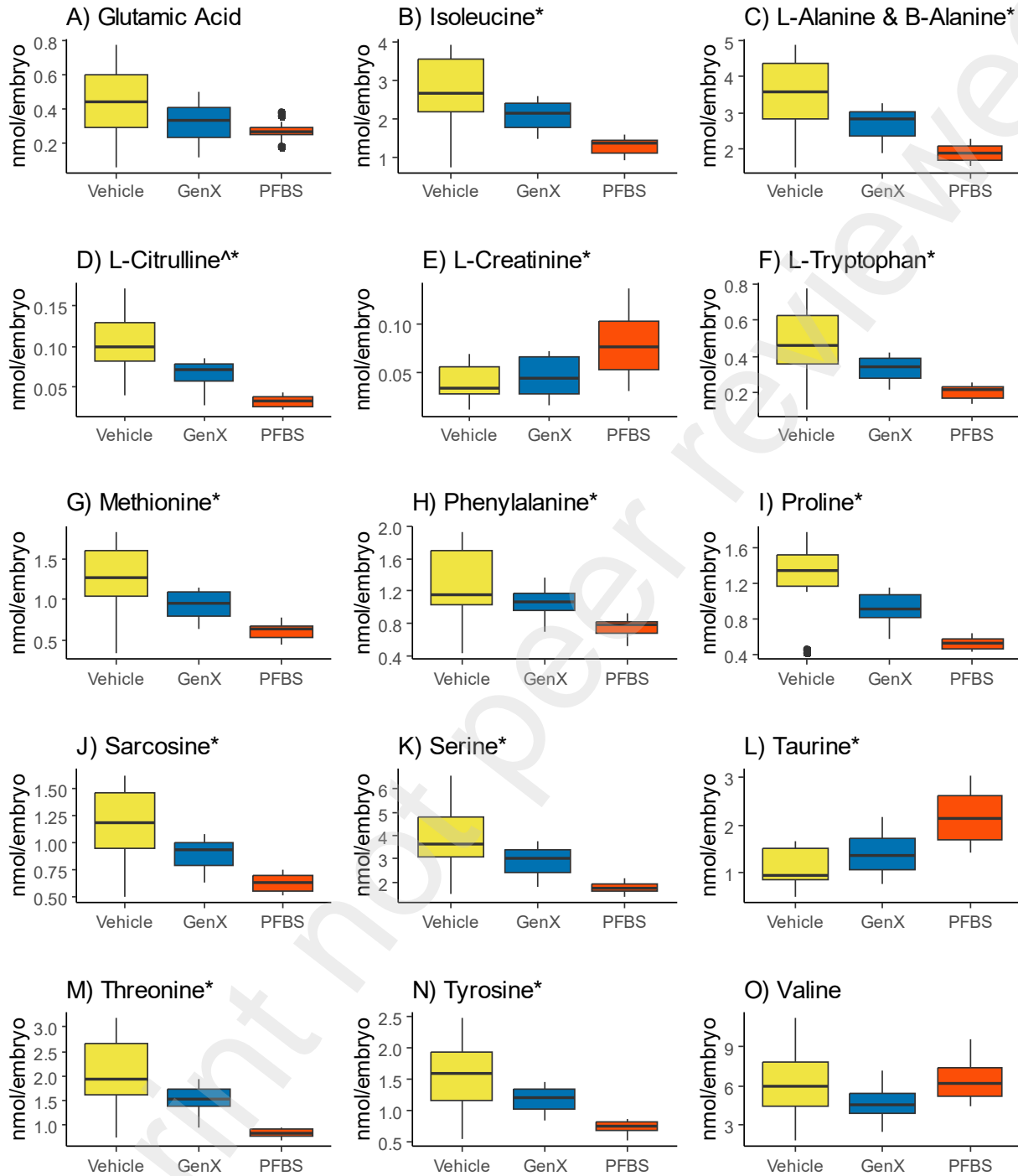


290 **Figure 2:** PFAS body burden of exposed to 100 μM PFBS or 80 μM GenX vs. vehicle groups. Each box plot represents
291 6 samples (each containing n=10 larvae) that were analyzed by isotope dilution mass spectrometry in triplicate.
292
293

294 **3.2 Targeted Metabolic Pathway Analysis**

295 In control and exposed larvae, 15 out of the 33 targeted metabolites were detected,
296 including glutamic acid, isoleucine, L-alanine and B-alanine, L-citrulline, L-creatinine, L-
297 tryptophan, methionine, phenylalanine, proline, sarcosine, serine, taurine, threonine, tyrosine, and
298 valine (**Figure 3; Table S5**). Concentrations ranged from <0.1 µg/embryo to 11.1 µg/embryo
299 across all samples and targeted metabolites.

300 In GenX exposed fish, one targeted metabolite, L-citrulline, was significantly decreased (p
301 < 0.05) (**Figure 3; Table S5**). In PFBS exposed fish, 11 targeted metabolites (isoleucine, L-alanine
302 and B-alanine, L-citrulline, L-tryptophan, methionine, phenylalanine, proline, sarcosine, serine,
303 threonine, and tyrosine) were significantly decreased (p < 0.05), while two metabolites were
304 significantly increased (L-creatinine and taurine) (p < 0.05).

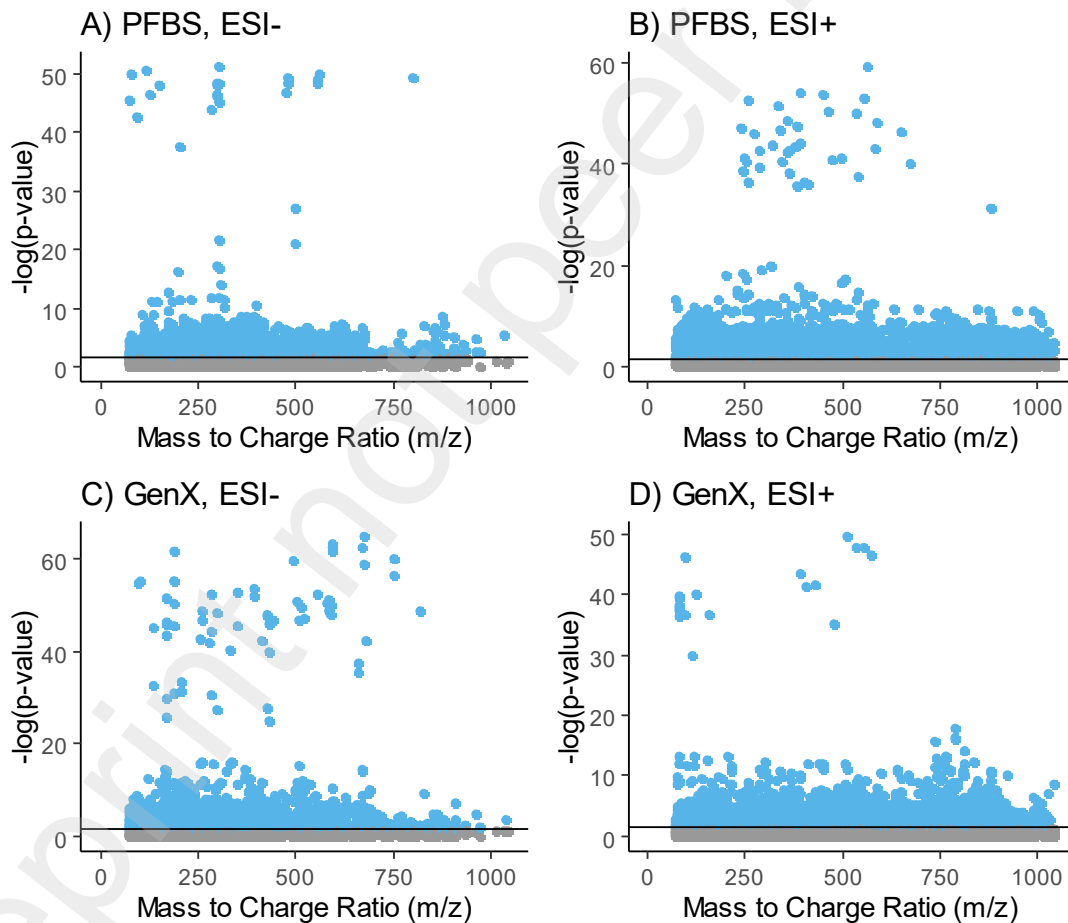


305
 306
 307
 308
 309
 310
 311

Figure 3: Targeted metabolites detected in GenX and PFBS exposed zebrafish embryos (each exposure group contained n = 6 groups of 10 larvae). * indicates significant alteration from the vehicle for PFBS exposure. ^ indicates significant alteration from the vehicle for GenX exposure.

312 **3.3 Untargeted Metabolomics: Pathway Analysis**

313 A metabolome-wide association study (MWAS) was performed to examine alterations in
314 the metabolome due to GenX or PFBS exposure in embryonic zebrafish. In GenX exposed
315 embryos, 831 of 4,554 detected features in the ESI- mode were significantly altered (FDR < 0.2),
316 and 630 of 8,517 detected features in ESI+ mode were significantly altered (FDR < 0.2) (**Figure**
317 **4**). From the PFBS exposure, 596 of 4,518 detected features in ESI- mode were significantly
318 altered (FDR < 0.2), and 1,897 of 8,532 detected features in ESI+ mode were significantly altered
319 (FDR < 0.2) (**Figure 4**).



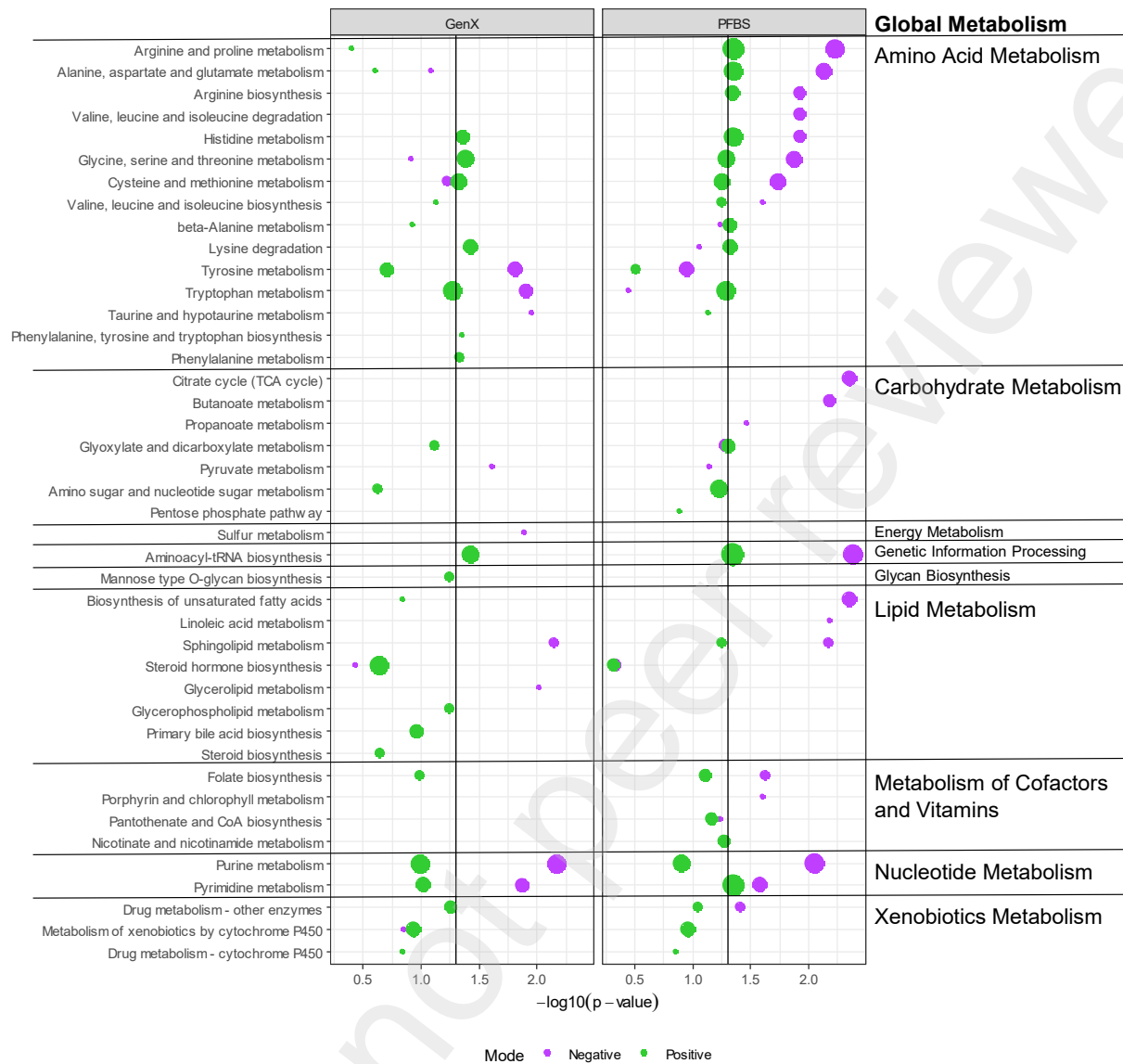
320

321 **Figure 4:** Manhattan plots for PFBS (a,b) and GenX (c,d) exposed larva, with ionization mode indicated as ESI+/-.
322 Each dot represents an individual feature. The horizontal black line indicated on each graph represents a FDR= 0.2;
323 all features (blue) above this line were significantly altered in the exposed groups in comparison to the vehicle group.
324

325 Features with $-\log_{10}(\text{p-value})$ above the 0.2 FDR threshold (**Figure 4**) were used in
326 pathway analysis to identify enriched pathways. Across both exposures, enriched pathways were
327 identified using 3 or more associated metabolites (features) for 42 metabolic pathways in total
328 (**Figure 5**). For the GenX exposure group, 32 total pathways were enriched with 18 unique to ESI+
329 mode, 13 overlapping between ESI+/- modes, and 1 unique to ESI- mode (**Table S6**). In the PFBS
330 exposure group, 34 total pathways were enriched with 5 unique to ESI+ mode, 21 overlapping
331 between ESI+/- modes, and 8 unique to ESI- mode (**Table S6**).

332 In the GenX exposed zebrafish, 17 out of 32 pathways were significantly ($p > 0.05$)
333 enriched in at least one ionization mode. In the PFBS exposed zebrafish, 21 out of 34 pathways
334 were significantly enriched in at least one of the two ionization modes. Enriched pathways were
335 grouped into nine different global metabolomic categories (**Table S6**): Amino acid metabolism,
336 Carbohydrate metabolism, Energy metabolism, Genetic information processing, Glycan
337 biosynthesis, Lipid metabolism, Metabolism of cofactors and vitamins, Nucleotide metabolism,
338 and Xenobiotics metabolism. Across all ionization modes and exposures, the global amino acid
339 metabolism contained the greatest number of enriched (15) and significantly enriched (14 out of
340 15 in at least one exposure group and ionization mode) pathways.

341



342

343 **Figure 5:** Enriched pathways from MetaboAnalyst pathway analysis, categorized by global metabolisms. The vertical
 344 black line represents $-\log_{10}(0.05)$ ($p\text{-value} = 0.05$). Values to the right of the line on each plot represent significantly
 345 altered pathways ($p\text{-value} > 0.05$).
 346

347

348

349

350

351 **4. DISCUSSION**

352 ***4.1 Body Burden***

353 Overall, the body burden data demonstrate a high uptake of each compound within the
354 embryos, as greater than 80% of the applied mass of each compound within exposure solutions
355 was absorbed by the larvae. However, the difference in body burden between each compound
356 provide insight into how readily each compound entered the zebrafish tissue. Despite a 73.7%
357 higher GenX exposure dose than PFBS (mg L⁻¹ comparison), GenX bioconcentration factor (BCF)
358 was lower than PFBS BCF. Results for both compounds demonstrate BCFs within a comparable
359 range to other known aqueous contaminants, with higher results indicating a more bioaccumulative
360 substance (Wassenaar et al., 2020; Winston H. Hickox, 2000). It is important to note that BCF can
361 vary greatly based on the population studied, and results from this study were used to determine
362 the difference between GenX and PFBS uptake into the zebrafish. The discrepancy between
363 exposure dose uptake and body burden suggests that PFBS more readily entered the tissue of
364 exposed zebrafish larvae than GenX or that GenX is more readily excreted from the larvae.

365 Additionally, a comparison between body burdens and metabolic pathway alterations
366 demonstrates that PFBS exposure resulted in a greater number of both significantly altered targeted
367 metabolites and significantly altered pathways despite a lower exposure dose and body burden.
368 The greater number of metabolic alterations observed in the PFBS group may indicate a higher
369 toxicity of PFBS exposure as compared to GenX exposure.

370

371 ***4.2 Biological Significance of Impacted Pathways***

372 *GenX Exposure:* Overall, GenX exposure demonstrated fewer impacted pathways or
373 metabolites than PFBS exposure, with only one targeted metabolite being significantly altered

374 from the vehicle group, and 32 altered pathways, of which 17 were significantly altered
375 (description of each of the significantly altered pathways can be found in **Table 1**). The
376 significantly altered pathways in the GenX exposed zebrafish fell into all nine global metabolisms:
377 Amino Acid Metabolism, Carbohydrate Metabolism, Energy Metabolism, Genetic Information
378 Processing, Lipid Metabolism, Metabolism of Cofactors and Vitamins, Nucleotide Metabolism,
379 and Xenobiotics Biodegradation and Metabolism (**Table S6**).

380

381

382

383

384

385

386

387
388

Table 1: Descriptions of significantly altered metabolic pathways, grouped into significantly altered metabolisms from the GenX exposure group only, both exposure groups, and the PFBS exposure group only.

Metabolic Pathway	Function	Reference
GenX		
Tryptophan Metabolism	Regulates energy homeostasis and aids in controlling inflammation, associated with aging.	(Sorgdrager et al., 2019)
Tyrosine Metabolism	Required for synthesis of proteins, used as an alternative energy source for cells. Used in the liver as a precursor to other metabolic processes.	(Nguyen et al., 2020)
Taurine and Hypotaurine Metabolism	Regulates processes in the central nervous system such as anti-oxidation and anti-inflammatory responses.	(Ichikawa et al., 2023)
Phenylalanine, Tyrosine, and Tryptophan Biosynthesis	Functions in the TCA cycle to help regulate energy expenditure and maintain energy homeostasis.	(Mishra et al., 2017)
Phenylalanine Metabolism	Functions in the TCA cycle to help regulate energy expenditure and maintain energy homeostasis. Acts as a precursor to the formation of important neurotransmitters.	(Mishra et al., 2017)
Pyruvate Metabolism	Involved in mitochondrial processes, end-product of glycolysis.	(Gray et al., 2014)
Sulfur Metabolism	Aids in utilizing sulfur-containing compounds, particularly the processing of cysteine and methionine amino acids.	(Griffith, 1987)
Glycerolipid Metabolism	Used to synthesize glycerolipids in the glycerolipid/free fatty acid cycle. Essential to the breakdown of lipids to maintain body temperature. Alterations can be associated with metabolic diseases such as metabolic syndrome.	(Prentki and Madiraju, 2008)
GenX and PFBS		
Cysteine and Methionine Metabolism	Regulation of metabolic processes, lipid metabolism, immune system, and digestion. Antioxidative effects through endogenous antioxidant enzyme activation and glutathione biosynthesis.	(Martínez et al., 2017)
Folate Biosynthesis	Involved in the synthesis of DNA, amino acids, and nucleotides. Alterations to this metabolism is associated with cellular damage, cardiovascular, and neurodegenerative disorders or diseases.	(Yao et al., 2022)
Glycine, Serine, and Threonine Metabolism	Required for synthesis of proteins, nucleic acids, and lipids. Biosynthesis of glycine and serine is essential to antioxidative properties of cells.	(Amelio et al., 2014)
Histidine Metabolism	Required for synthesis of proteins used in the small intestine and liver, incorporated into amino acids essential to muscle and brain tissue.	(Moro et al., 2020)
Lysine Degradation	Involved in neurometabolic and mitochondrial processes.	(Leandro and Houten, 2020)
Aminoacyl-tRNA Biosynthesis	Required for protein synthesis and the production of transfer RNAs. Function in the immune system as regulators and signaling molecules for immune cell development and regulators for biological processes in immune cells.	(Nie et al., 2019)
Purine Metabolism	Structural component of DNA and RNA, functional component of many metabolic processes. Disruptions to the purine metabolism is associated with dysregulation of the nervous system in humans.	(Ren et al., 2018)
Pyrimidine Metabolism	Essential to the synthesis of DNA and RNA, used for cell growth, development, signaling, and energy.	(Wang et al., 2021)
Sphingolipid Metabolism	Found in nervous tissue, sphingolipids regulate cell function and accumulate during periods of oxidative stress.	(Doccini et al., 2022)
PFBS		
Alanine, Aspartate and Glutamate Metabolism	Derived from the Citric Acid (TCA) Cycle and is a precursor for other amino acids and several important metabolic processes.	(Reitzer, 2004)

Arginine and Proline Metabolism	Proline acts as a precursor to arginine, proline metabolism acts as metabolic regulation pathway to account for deficiencies in associated pathways. Arginine and proline play roles in nitrogen cycling in organisms.	(Phang, 2017)
Arginine Biosynthesis	Precursor to other amino acids, plays a role in protein synthesis.	(Charlier and Glansdorff, 2004)
Beta-Alanine Metabolism	Involved in glycolysis through combination with histidine to form carnosine.	(Vaughan et al., 2014)
Biosynthesis of Unsaturated Fatty Acids	Formed in the liver for cell walls, membranes, and protein modification. Used for energy storage and a precursor to other metabolites.	(Beld et al., 2015)
Butanoate Metabolism	Involved in the biosynthesis of polysaccharides in the gut lining, and play an essential role in intestinal immune responses.	(Liu et al., 2022; Wang et al., 2020)
Citrate Cycle (TCA Cycle)	Main energy metabolism within organisms. Oxidizes fats, proteins, and carbohydrates to product energy.	(Martínez-Reyes and Chandel, 2020)
Glyoxylate and Dicarboxylate Metabolism	Utilizes fatty acids and acetate to produce energy as an alternative pathway to the TCA Cycle. Alterations associated with oxidative stress.	(Ahn et al., 2016)
Valine, Leucine, and Isoleucine Biosynthesis	Branched chain amino acids that play a role in the energy metabolism of zebrafish. Alterations can be associated with oxidative stress and organ damage.	(Meng et al., 2022)
Linoleic Acid Metabolism	Fatty acid metabolism to process dietary linoleic acid.	(Banni, 2002)
Porphyrin and Chlorophyll Metabolism	Essential to the formation of hemoglobin.	(Zhou et al., 2020)
Drug Metabolism – Other Enzymes	Enzymes that play a role in drug absorption and processing within the zebrafish.	(Hung et al., 2012)

389

390 From the significantly altered metabolisms resulting from GenX exposure, there are several
391 common metabolic systems and biological processes that were affected by multiple metabolic
392 pathways, including protein synthesis, cellular signaling, growth and development, antioxidative
393 properties, energy processes, and immune and nervous system regulation. Disruptions to these
394 pathways, individually or in combination with each other, could have effects on many organ
395 systems and biological processes, such as response to oxidative stress or disease, liver
396 functionality, and maintenance of metabolic processes through cellular development. In humans,
397 alterations to several of the identified significantly altered pathways are associated with severe

398 diseases, including neurodegenerative diseases such as Alzheimer's and Parkinson's Diseases
399 (purine metabolism), cancerous tumor growth (glycine, serine, and threonine metabolism), and
400 Metabolic Syndrome (association with weight gain, Type 2 Diabetes, obesity (alanine, aspartate,
401 and glutamate metabolism; cysteine and methionine metabolism; glycerolipid metabolism;
402 glycine, serine, and threonine metabolism) (Amelio et al., 2014; Ansoleaga et al., 2015; Chen et
403 al., 2022; Garcia-Esparcia et al., 2015; Li et al., 2020). A previous study of zebrafish embryonic
404 and larval exposure to PFOS and other toxicants found that alterations to the alanine and proline
405 metabolism indicated oxidative stress and imbalances in homeostasis and energy metabolisms
406 (Ortiz-Villanueva et al., 2018). A different study discovered that alterations to the methionine
407 metabolism had adverse effects on liver formation during zebrafish development (Liu et al., 2016).
408 Our observed perturbations to metabolic pathways suggest metabolic dysregulation in the
409 zebrafish embryos, with potential physiological implications ranging from liver development to
410 tissue maintenance.

411 In addition to the functions listed in **Table 1**, several significantly altered pathways are
412 associated with essential amino acids, or amino acids that must be consumed in animals' diets for
413 proper development and maintenance of cells, tissues, and in bodily functions. From the list of
414 impacted metabolic pathways in **Table 1**, the following are nutritionally essential amino acids:
415 cysteine, histidine, lysine, methionine, threonine, tryptophan, and tyrosine (Watts and D'Abramo,
416 2021). Disruption of essential amino acid pathways for processing and utilizing amino acids in the
417 body can have effects on growth, survival, reproduction, and overall health of the animal, which
418 indicates that metabolic pathways essential to development and health of the zebrafish have been
419 impacted by exposure to GenX in this study (Watts and D'Abramo, 2021). In combination with
420 the impacts of altered metabolisms that affect cellular development and growth, protein synthesis

421 and muscular maintenance, and metabolic homeostasis, exposure to GenX during the embryonic
422 and larval stages may have impacts on the proper development and growth of the zebrafish.

423 *PFBS Exposure:* Despite lower body burden in the PFBS exposure group, PFBS exposure
424 resulted in a greater number of significantly altered pathways and targeted metabolites. Significant
425 overlaps exist between significantly altered pathways from the GenX exposure group and PFBS
426 exposure group; there were 12 significantly altered pathways uniquely in the PFBS exposure group
427 in addition to the 9 shared with the GenX exposed zebrafish (**Table 1**). Significantly altered
428 pathways for the PFBS exposed zebrafish (including overlap with the GenX exposed zebrafish)
429 were in 8 global metabolisms: amino acid metabolism, carbohydrate metabolism, genetic
430 information processing, lipid metabolism, metabolism of cofactors and vitamins, nucleotide
431 metabolism, and xenobiotics biodegradation and metabolism (**Table S6**). PFBS exposure resulted
432 in 12 additional significantly altered targeted metabolites with the only overlap with the GenX
433 exposure group being L-Citrulline, which was significantly downregulated in both exposures
434 (**Table S5**).

435 Several metabolic pathways affected in the PFBS exposure group further contribute to
436 Metabolic Syndrome, including alanine, aspartate, and glutamate metabolism, arginine
437 biosynthesis, TCA Cycle, and valine, leucine, and isoleucine biosynthesis (Chen et al., 2022).
438 Several metabolic pathway alterations overlap with significantly altered metabolites from targeted
439 analysis, including alterations to levels of alanine, methionine, phenylalanine, proline, serine,
440 threonine, tryptophan, tyrosine, and valine. There are also several additional functions impacted
441 by the pathways that were altered uniquely in the PFBS exposure group, specifically the alterations
442 observed in the energy/carbohydrate metabolism and its related amino acid pathways (TCA Cycle;
443 glyoxylate and dicarboxylate metabolism; alanine, aspartate, and glutamate metabolism) and

444 effects on the lipid metabolism through use of fatty acids. Previous zebrafish toxicology studies
445 indicate that alterations to the lipid (fatty acid) metabolism affects lipid creation and utilization in
446 the liver, and disturbances to the carbohydrate metabolism may indicate insulin resistance,
447 associated with diabetes (Du et al., 2016; Teng et al., 2018). In humans, diabetes can be a result of
448 Metabolic Syndrome, so alterations to the amino acid metabolisms affected in both exposures as
449 well as the alterations to the lipid metabolism observed in the PFBS exposure group may indicate
450 chronic perturbations to organ functions and bodily regulation (Chen et al., 2022).

451 It is also important to consider the implications of these results for human health, especially
452 if contamination of water with replacement PFAS compounds such as PFBS and GenX becomes
453 more widespread. The primary route of exposure to PFAS is through consumption of contaminated
454 food and drinking water, and studies show that consumption of freshwater fish contaminated with
455 PFAS can cause a direct increase in blood serum levels of PFAS in humans (Barbo et al., 2023).
456 Considering the high uptake of GenX and PFBS by zebrafish in this study, ingestion of these
457 compounds could pose dangers to human health through an increase in blood serum levels of GenX
458 and PFBS. Not only would consumption of contaminated fish pose risks to the consumer, but the
459 effects observed in this study would likely be most relevant to *in utero* exposure to PFAS. Fetal
460 exposure to PFAS during pregnancy can cause the fetus to be at higher risk for metabolic
461 syndrome, which is consistent with metabolites altered in this study, as well as thyroid dysfunction
462 and kidney disease (Blake and Fenton, 2020). Using zebrafish as a toxicological model for both
463 fish uptake of novel PFAS as well as potential human health effects demonstrates that adverse
464 health effects are possible in vertebrate populations exposed to GenX and PFBS.

465

466 **4.3 Study Limitations**

467 The single analyte exposure paradigm used in this study was different from
468 environmentally relevant exposures to these chemicals. Across the US, freshwater fish are
469 chronically exposed to combinations of PFAS compounds (concentrations depending on
470 geographic location) at lower exposure doses than used in this study (Barbo et al., 2023). Even in
471 the Cape Fear River, a site of known GenX contamination due to an upstream chemical plant, the
472 maximum concentration of GenX in a water quality study was $4.56 \mu\text{g L}^{-1}$, with an average of 630
473 ng L^{-1} , indicating that the 49.5 mg L^{-1} exposure dose used in this study was an order of magnitude
474 higher than GenX concentrations found in surface waters across the US (Sun et al., 2016).
475 Although our exposure doses may not be representative of chronic exposure levels, the results are
476 relevant to embryonic or early life exposure to GenX or PFBS. Therefore, developmental
477 metabolome alterations from this study could be detected to a lesser extent at lower exposure doses
478 observed in aquatic environments.

479 This study also included several limiting factors that do not allow for further comparison
480 or analysis of the toxicity and metabolic effects of GenX and PFBS. Firstly, the exposure doses
481 used were based on mortality and/or physiological alterations that occurred due to exposure;
482 however, the concentrations used in exposure were not equal, nor were there multiple
483 concentrations to compare the level of toxicity of each compound to the embryos. For example, it
484 would be useful to assess exposure doses for each compound leading to a similar magnitude or
485 number of metabolic, as we observed a greater number of metabolic alterations in the PFBS
486 exposure group with a different exposure dose. Additionally, this study focused only on the
487 individual effects of each compound; future studies would ideally assess exposure to a mixture of
488 PFBS and GenX as well as other short-chain PFAS. While there are limitations to this study,
489 including unequal exposure doses and lack of multiple dosing concentrations or additional dosing

490 compounds, the study as a whole provides a basis for future studies that can provide additional
491 insight into the toxicity of GenX and PFBS to zebrafish and other vertebrates.

492
493 **CONCLUSIONS**

494 Overall, this study provides an understanding of metabolic alterations induced by exposure
495 to GenX or PFBS in zebrafish embryos. These exposures resulted in no physiological deformities
496 or excess mortality of the zebrafish over a 6-day exposure period. The body burden assessment
497 indicated that either PFBS is more readily absorbed in zebrafish larvae compared to GenX or GenX
498 is more readily excreted from zebrafish larvae. Targeted and untargeted metabolomics were used
499 to understand alterations to the zebrafish metabolome induced by exposure to GenX and PFBS,
500 and PFBS exposure induced a greater number of significantly altered metabolites and metabolic
501 pathways than GenX exposure. Altered metabolic pathways in both groups indicate altered
502 responses to oxidative stress, dysregulation of energy homeostasis and growth-related metabolites,
503 and perturbations to lipid creation in the liver. These results indicate that developmental exposure
504 to GenX or PFBS may disrupt cellular development and metabolic regulation in multiple organ
505 systems. Using zebrafish as a toxicological model, the results from this study demonstrate the
506 metabolic effects of early-life exposure to GenX and PFBS. Further studies on combined effects
507 of GenX, PFBS, and other short-chain replacement PFAS will provide greater detail on potential
508 health effects of ingestion of these chemicals for both humans and other animals.

509
510
511 **ACKNOWLEDGEMENTS**

512 FD was supported by the Brown University School of Engineering DiMase Summer Internship
513 Fellowship. KM was supported by NIEHS K01 ES035398 (PI) and R01 ES032386. The Thermo
514 LC-Orbitrap MS was partially funded by NSF Major Research Instrumentation (MRI) award

515 CBET-1919870 to KDP (PI) and JSP (Co-I). SEP was supported by a Ruth L. Kirschstein
516 Predoctoral Individual National Research Service Award (NRSA; F31HL156460) from the
517 NHLBI and was previously supported by the Brown University Environmental Pathology Training
518 Grant (T32ES007272-26). JSP was supported by a NIEHS Outstanding New Environmental
519 Scientist (ONES) award (R01ES030109).

520

521 **AUTHOR CONTRIBUTIONS**

522 FD: Conceptualization, Investigation, Formal Analysis, Original Draft, Reviewing and Editing.

523 SP: Investigation, Methodology, Reviewing and Editing. KDP: Conceptualization, Resources,

524 Reviewing and Editing. JP: Methodology, Resources, Supervision, Reviewing and Editing. KEM:

525 Conceptualization, Methodology, Validation, Data Curation, Supervision, Reviewing and Editing.

526

527

528 **REFERENCES**

529 Ahn S, Jung J, Jang IA, Madsen EL, Park W. Role of Glyoxylate Shunt in Oxidative Stress
530 Response. *J Biol Chem* 2016; 291: 11928-38.

531 Amelio I, Cutruzzolá F, Antonov A, Agostini M, Melino G. Serine and glycine metabolism in
532 cancer. *Trends Biochem Sci* 2014; 39: 191-8.

533 Ansoleaga B, Jové M, Schlüter A, Garcia-Esparcia P, Moreno J, Pujol A, et al. Deregulation of
534 purine metabolism in Alzheimer's disease. *Neurobiology of Aging* 2015; 36: 68-80.

535 Banni S. Conjugated linoleic acid metabolism. *Current Opinion in Lipidology* 2002; 13.

536 Bao Y, Deng S, Jiang X, Qu Y, He Y, Liu L, et al. Degradation of PFOA Substitute: GenX
537 (HFPO-DA Ammonium Salt): Oxidation with UV/Persulfate or Reduction with
538 UV/Sulfite? *Environmental Science & Technology* 2018; 52: 11728-11734.

539 Barbo N, Stoiber T, Naidenko OV, Andrews DQ. Locally caught freshwater fish across the
540 United States are likely a significant source of exposure to PFOS and other perfluorinated
541 compounds. *Environmental Research* 2023; 220: 115165.

542 Beld J, Lee DJ, Burkart MD. Fatty acid biosynthesis revisited: structure elucidation and
543 metabolic engineering. *Mol Biosyst* 2015; 11: 38-59.

544 Blake BE, Cope HA, Hall SM, Keys RD, Mahler BW, McCord J, et al. Evaluation of Maternal,
545 Embryo, and Placental Effects in CD-1 Mice following Gestational Exposure to
546 Perfluorooctanoic Acid (PFOA) or Hexafluoropropylene Oxide Dimer Acid (HFPO-DA
547 or GenX). *Environ Health Perspect* 2020; 128: 27006.

548 Blake BE, Fenton SE. Early life exposure to per- and polyfluoroalkyl substances (PFAS) and
549 latent health outcomes: A review including the placenta as a target tissue and possible
550 driver of peri- and postnatal effects. *Toxicology* 2020; 443: 152565.

551 Charlier D, Glansdorff N. Biosynthesis of Arginine and Polyamines. *EcoSal Plus* 2004; 1.

552 Chen F, Wei C, Chen Q, Zhang J, Wang L, Zhou Z, et al. Internal concentrations of
553 perfluorobutane sulfonate (PFBS) comparable to those of perfluorooctane sulfonate
554 (PFOS) induce reproductive toxicity in *Caenorhabditis elegans*. *Ecotoxicol Environ Saf*
555 2018; 158: 223-229.

556 Chen M, Yang Z, Gan H, Wang Y, Li C, Gao Y. Investigation into potential mechanisms of
557 metabolic syndrome by integrative analysis of metabolomics and proteomics. *PLOS ONE*
558 2022; 17: e0270593.

559 Cheng J, Lv S, Nie S, Liu J, Tong S, Kang N, et al. Chronic perfluorooctane sulfonate (PFOS)
560 exposure induces hepatic steatosis in zebrafish. *Aquat Toxicol* 2016; 176: 45-52.

561 Chong J, Xia J. *MetaboAnalystR*: an R package for flexible and reproducible analysis of
562 metabolomics data. *Bioinformatics* 2018; 34: 4313-4314.

563 da Silva KM, Iturrospe E, Bars C, Knapen D, Van Cruchten S, Covaci A, et al. Mass
564 Spectrometry-Based Zebrafish Toxicometabolomics: A Review of Analytical and Data
565 Quality Challenges. *Metabolites* 2021; 11.

566 Doccini S, Marchese M, Morani F, Gammaldi N, Mero S, Pezzini F, et al. Lysosomal Proteomics
567 Links Disturbances in Lipid Homeostasis and Sphingolipid Metabolism to CLN5
568 Disease. *Cells*. 11, 2022.

569 Du Z, Zhang Y, Wang G, Peng J, Wang Z, Gao S. TPhP exposure disturbs carbohydrate
570 metabolism, lipid metabolism, and the DNA damage repair system in zebrafish liver.
571 *Scientific Reports* 2016; 6: 21827.

572 Gaballah S, Swank A, Sobus Jon R, Howey Xia M, Schmid J, Catron T, et al. Evaluation of
573 Developmental Toxicity, Developmental Neurotoxicity, and Tissue Dose in Zebrafish
574 Exposed to GenX and Other PFAS. *Environmental Health Perspectives* 2020; 128:
575 047005.

576 Garcia-Esparcia P, Hernández-Ortega K, Ansoleaga B, Carmona M, Ferrer I. Purine metabolism
577 gene deregulation in Parkinson's disease. *Neuropathology and Applied Neurobiology*
578 2015; 41: 926-940.

579 Gebreab KY, Eeza MNH, Bai T, Zuberi Z, Matysik J, O'Shea KE, et al. Comparative
580 toxicometabolomics of perfluorooctanoic acid (PFOA) and next-generation
581 perfluoroalkyl substances. *Environmental Pollution* 2020; 265: 114928.

582 Glüge J, Scheringer M, Cousins IT, DeWitt JC, Goldenman G, Herzke D, et al. An overview of
583 the uses of per- and polyfluoroalkyl substances (PFAS). *Environmental Science:
584 Processes & Impacts* 2020; 22: 2345-2373.

585 Gray LR, Tompkins SC, Taylor EB. Regulation of pyruvate metabolism and human disease.
586 *Cellular and Molecular Life Sciences* 2014; 71: 2577-2604.

587 Griffith OW. Mammalian sulfur amino acid metabolism: An overview. *Methods in Enzymology*.
588 143. Academic Press, 1987, pp. 366-376.

589 Henn K, Braunbeck T. Dechoriation as a tool to improve the fish embryo toxicity test (FET)
590 with the zebrafish (*Danio rerio*). *Comp Biochem Physiol C Toxicol Pharmacol* 2011; 153:
591 91-8.

592

593 Howe K, Clark MD, Torroja CF, Torrance J, Berthelot C, Muffato M, et al. The zebrafish
594 reference genome sequence and its relationship to the human genome. *Nature* 2013; 496:
595 498-503.

596 Hu C, Huang Z, Sun B, Liu M, Tang L, Chen L. Metabolomic profiles in zebrafish larvae
597 following probiotic and perfluorobutanesulfonate coexposure. *Environmental Research*
598 2022; 204: 112380.

599 Hung MW, Zhang ZJ, Li S, Lei B, Yuan S, Cui GZ, et al. From Omics to Drug Metabolism and
600 High Content Screen of Natural Product in Zebrafish: A New Model for Discovery of
601 Neuroactive Compound. *Evidence-Based Complementary and Alternative Medicine*
602 2012; 2012: 605303.

603 Ichikawa S, Abe R, Fujimoto H, Higashi K, Zang L, Nakayama H, et al. Paraburkholderia sabiae
604 administration alters zebrafish anxiety-like behavior via gut microbial taurine
605 metabolism. *Frontiers in Microbiology* 2023; 14.

606 Janousek RM, Mayer J, Knepper TP. Is the phase-out of long-chain PFASs measurable as
607 fingerprint in a defined area? Comparison of global PFAS concentrations and a
608 monitoring study performed in Hesse, Germany from 2014 to 2018. *TrAC Trends in*
609 *Analytical Chemistry* 2019; 120: 115393.

610 Kabadi SV, Fisher JW, Doerge DR, Mehta D, Aungst J, Rice P. Characterizing biopersistence
611 potential of the metabolite 5:3 fluorotelomer carboxylic acid after repeated oral exposure
612 to the 6:2 fluorotelomer alcohol. *Toxicology and Applied Pharmacology* 2020; 388:
613 114878.

614 Kanehisa M, Goto S. KEGG: Kyoto Encyclopedia of Genes and Genomes. *Nucleic Acids*
615 *Research* 2000; 28: 27-30.

616 Kossack ME, Manz KE, Martin NR, Pennell KD, Plavicki J. Environmentally relevant uptake,
617 elimination, and metabolic changes following early embryonic exposure to 2,3,7,8-
618 tetrachlorodibenzo-p-dioxin in zebrafish. *Chemosphere* 2023; 310: 136723.

619 Kucharzyk KH, Darlington R, Benotti M, Deeb R, Hawley E. Novel treatment technologies for
620 PFAS compounds: A critical review. *Journal of Environmental Management* 2017; 204:
621 757-764.

622 Lai KP, Gong Z, Tse WKF. Zebrafish as the toxicant screening model: Transgenic and omics
623 approaches. *Aquatic Toxicology* 2021; 234: 105813.

624 Leandro J, Houten SM. The lysine degradation pathway: Subcellular compartmentalization and
625 enzyme deficiencies. *Molecular Genetics and Metabolism* 2020; 131: 14-22.

626 Li S, Park Y, Duraisingham S, Strobel FH, Khan N, Soltow QA, et al. Predicting Network
627 Activity from High Throughput Metabolomics. *PLOS Computational Biology* 2013; 9:
628 e1003123.

629 Li Y, Wang Y, Zhuang Y, Zhang P, Chen S, Asakawa T, et al. Serum Metabolomic Profiles
630 Associated With Untreated Metabolic Syndrome Patients in the Chinese Population.
631 *Clinical and Translational Science* 2020; 13: 1271-1278.

632 Liu C, Hua H, Guo Y, Qian H, Liu J, Cheng Y. Study on the hepatoprotective effect of
633 *Sporidiobolus pararoseus* polysaccharides under the “gut microbiome-amino acids
634 metabolism” network. *Food Bioscience* 2022; 49: 101928.

635 Liu LY, Alexa K, Cortes M, Schatzman-Bone S, Kim AJ, Mukhopadhyay B, et al. Cannabinoid
636 receptor signaling regulates liver development and metabolism. *Development* 2016; 143:
637 609-622.

- 638 Martínez Y, Li X, Liu G, Bin P, Yan W, Más D, et al. The role of methionine on metabolism,
639 oxidative stress, and diseases. *Amino Acids* 2017; 49: 2091-2098.
- 640 Martínez-Reyes I, Chandel NS. Mitochondrial TCA cycle metabolites control physiology and
641 disease. *Nature Communications* 2020; 11: 102.
- 642 Meng Z, Cui J, Liu L, Yang C, Bao X, Wang J, et al. Toxicity effects of chlorantraniliprole in
643 zebrafish (*Danio rerio*) involving in liver function and metabolic phenotype. *Pesticide*
644 *Biochemistry and Physiology* 2022; 187: 105194.
- 645 Min EK, Lee H, Sung EJ, Seo SW, Song M, Wang S, et al. Integrative multi-omics reveals
646 analogous developmental neurotoxicity mechanisms between perfluorobutanesulfonic
647 acid and perfluorooctanesulfonic acid in zebrafish. *J Hazard Mater* 2023; 457: 131714.
- 648 Mishra P, Gong Z, Kelly BC. Assessing biological effects of fluoxetine in developing zebrafish
649 embryos using gas chromatography-mass spectrometry based metabolomics.
650 *Chemosphere* 2017; 188: 157-167.
- 651 Mishra P, Singh U, Pandey CM, Mishra P, Pandey G. Application of student's t-test, analysis of
652 variance, and covariance. *Ann Card Anaesth* 2019; 22: 407-411.
- 653 Moro J, Tomé D, Schmidely P, Demersay T-C, Azzout-Marniche D. Histidine: A Systematic
654 Review on Metabolism and Physiological Effects in Human and Different Animal
655 Species. *Nutrients*. 12, 2020.
- 656 Newsted JL, Beach SA, Gallagher SP, Giesy JP. Acute and chronic effects of perfluorobutane
657 sulfonate (PFBS) on the mallard and northern bobwhite quail. *Arch Environ Contam*
658 *Toxicol* 2008; 54: 535-45.
- 659 Nguyen TN, Nguyen HQ, Le D-H. Unveiling prognostics biomarkers of tyrosine metabolism
660 reprogramming in liver cancer by cross-platform gene expression analyses. *PLOS ONE*
661 2020; 15: e0229276.
- 662 Nie A, Sun B, Fu Z, Yu D. Roles of aminoacyl-tRNA synthetases in immune regulation and
663 immune diseases. *Cell Death & Disease* 2019; 10: 901.
- 664 Ortiz-Villanueva E, Jaumot J, Martínez R, Navarro-Martín L, Piña B, Tauler R. Assessment of
665 endocrine disruptors effects on zebrafish (*Danio rerio*) embryos by untargeted LC-HRMS
666 metabolomic analysis. *Science of The Total Environment* 2018; 635: 156-166.
- 667 Phang JM. Proline Metabolism in Cell Regulation and Cancer Biology: Recent Advances and
668 Hypotheses. *Antioxidants & Redox Signaling* 2017; 30: 635-649.
- 669 Podder A, Sadmani AHMA, Reinhart D, Chang N-B, Goel R. Per and poly-fluoroalkyl
670 substances (PFAS) as a contaminant of emerging concern in surface water: A
671 transboundary review of their occurrences and toxicity effects. *Journal of Hazardous*
672 *Materials* 2021; 419: 126361.
- 673 Prentki M, Madiraju SRM. Glycerolipid Metabolism and Signaling in Health and Disease.
674 *Endocrine Reviews* 2008; 29: 647-676.
- 675 Rayne S, Forest K. Perfluoroalkyl sulfonic and carboxylic acids: A critical review of
676 physicochemical properties, levels and patterns in waters and wastewaters, and treatment
677 methods. *Journal of Environmental Science and Health, Part A* 2009; 44: 1145-1199.
- 678 Reitzer L. Biosynthesis of Glutamate, Aspartate, Asparagine, L-Alanine, and D-Alanine. *EcoSal*
679 *Plus* 2004; 1.
- 680 Ren X, Zhang H, Geng N, Xing L, Zhao Y, Wang F, et al. Developmental and metabolic
681 responses of zebrafish (*Danio rerio*) embryos and larvae to short-chain chlorinated
682 paraffins (SCCPs) exposure. *Science of The Total Environment* 2018; 622-623: 214-221.

683 Rericha Y, Cao D, Truong L, Simonich MT, Field JA, Tanguay RL. Sulfonamide functional
684 head on short-chain perfluorinated substance drives developmental toxicity. *iScience*
685 2022; 25: 103789.

686 Sorgdrager FJH, Naudé PJW, Kema IP, Nollen EA, Deyn PPD. Tryptophan Metabolism in
687 Inflammaging: From Biomarker to Therapeutic Target. *Frontiers in Immunology* 2019;
688 10.

689 Sukardi H, Ung CY, Gong Z, Lam SH. Incorporating Zebrafish Omics into Chemical Biology
690 and Toxicology. *Zebrafish* 2010; 7: 41-52.

691 Sun M, Arevalo E, Strynar M, Lindstrom A, Richardson M, Kearns B, et al. Legacy and
692 Emerging Perfluoroalkyl Substances Are Important Drinking Water Contaminants in the
693 Cape Fear River Watershed of North Carolina. *Environmental Science & Technology*
694 Letters 2016; 3: 415-419.

695 Sunderland EM, Hu XC, Dassuncao C, Tokranov AK, Wagner CC, Allen JG. A review of the
696 pathways of human exposure to poly- and perfluoroalkyl substances (PFASs) and present
697 understanding of health effects. *Journal of Exposure Science & Environmental*
698 *Epidemiology* 2019; 29: 131-147.

699 Taniyasu S, Kannan K, Horii Y, Hanari N, Yamashita N. A Survey of Perfluorooctane Sulfonate
700 and Related Perfluorinated Organic Compounds in Water, Fish, Birds, and Humans from
701 Japan. *Environmental Science & Technology* 2003; 37: 2634-2639.

702 Teng M, Zhu W, Wang D, Yan J, Qi S, Song M, et al. Acute exposure of zebrafish embryo
703 (*Danio rerio*) to flutolanil reveals its developmental mechanism of toxicity via disrupting
704 the thyroid system and metabolism. *Environmental Pollution* 2018; 242: 1157-1165.

705 USEPA. Draft Method 1633: Analysis of Per- and Polyfluoroalkyl Substances (PFAS) in
706 Aqueous, Solid, Biosolids, and Tissue Samples by LC-MS/MS. In: Agency EP, editor,
707 2021.

708 USEPA. Fifth Unregulated Contaminant Monitoring Rule, 2022, pp. 2022 drinking water
709 advisory levels for PFOA, PFOS, GenX, and PFBS.

710 USEPA. PFAS National Primary Drinking Water Regulation Rulemaking. In: Agency EP,
711 editor. 88. USEPA, 2023.

712 Vaughan RA, Gannon NP, Garcia-Smith R, Licon-Munoz Y, Barberena MA, Bisoffi M, et al. β -
713 alanine suppresses malignant breast epithelial cell aggressiveness through alterations in
714 metabolism and cellular acidity in vitro. *Molecular Cancer* 2014; 13: 14.

715 Wallington TJ, Andersen MPS, Nielsen OJ. The case for a more precise definition of regulated
716 PFAS. *Environmental Science: Processes & Impacts* 2021; 23: 1834-1838.

717 Wang D, Guo S, He H, Gong L, Cui H. Gut Microbiome and Serum Metabolome Analyses
718 Identify Unsaturated Fatty Acids and Butanoate Metabolism Induced by Gut Microbiota
719 in Patients With Chronic Spontaneous Urticaria. *Frontiers in Cellular and Infection*
720 *Microbiology* 2020; 10.

721 Wang P, Lu Y, Wang T, Zhu Z, Li Q, Meng J, et al. Coupled production and emission of short
722 chain perfluoroalkyl acids from a fast developing fluorochemical industry: Evidence from
723 yearly and seasonal monitoring in Daling River Basin, China. *Environmental Pollution*
724 2016; 218: 1234-1244.

725 Wang W, Cui J, Ma H, Lu W, Huang J. Targeting Pyrimidine Metabolism in the Era of Precision
726 Cancer Medicine. *Frontiers in Oncology* 2021; 11.

727 Wassenaar PNH, Verbruggen EMJ, Cieraad E, Peijnenburg WJGM, Vijver MG. Variability in fish
728 bioconcentration factors: Influences of study design and consequences for regulation.
729 Chemosphere 2020; 239: 124731.
730
731 Watts SA, D'Abramo LR. Standardized Reference Diets for Zebrafish: Addressing Nutritional
732 Control in Experimental Methodology. Annu Rev Nutr 2021; 41: 511-527.
733 Westerfield M. The Zebrafish Book : A Guide for the Laboratory Use of Zebrafish.
734 http://zfin.org/zf_info/zfbook/zfbk.html 2000.
735 Winston H. Hickox JED, Ph.D. Technical Support Document for Exposure Assessment and
736 Stochastic Analysis. In: Agency CEP, editor. California Environmental Protection
737 Agency, 2000.
738
739 Yao W, Chen J, Lin Z, Wang N, Wang A, Wang B, et al. Scopoletin Induced Metabolomic
740 Profile Disturbances in Zebrafish Embryos. Metabolites. 12, 2022.
741 Zhou Z-Y, Zhao W-R, Xiao Y, Zhang J, Tang J-Y, Lee SM-Y. Mechanism Study of the
742 Protective Effects of Sodium Tanshinone IIA Sulfonate Against Atorvastatin-Induced
743 Cerebral Hemorrhage in Zebrafish: Transcriptome Analysis. Frontiers in Pharmacology
744 2020; 11.
745
746
747
748